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2 **Evaluating Health Risk Indicators for PTE Exposure in the Food**
3 **Chain: evidence from a Thallium Mine Area**

4 Feng Jiang^{1,2}, Bozhi Ren^{1,2}, Andrew S Hursthouse^{1,3} & Renjian Deng^{1,2}

5 **1 Hunan Provincial Key Laboratory of Shale Gas Resource Exploitation, Xiangtan 411201, China**

6 **2 School of Civil Engineering, Hunan University of Science and Technology, Xiangtan 411201, China**

7 **3 Computing Engineering & Physical Sciences, University of the West of Scotland, Paisley PA1 2BE,**

8 **UK**

9 **Abstract:** Thallium (Tl) has a high relative toxicity and is easily taken up by plants but little is known about
10 wider relationship with co-contaminants and in typical domestic food crops. We evaluated the extent of
11 contamination, evidence for bioaccumulation in typical food crops (Chinese cabbage, green cabbage, chili,
12 carrot, corn and rice) and subsequent contribution to health risks for 7 elements (Tl, As, Cd, Pb, Ni, Cu, Zn)
13 associated with soil contamination in the local soils of a major Tl mine in Guizhou, southwest China.
14 Derivation of relevant risk indicators from the bioconcentration factor (BCF), comprehensive crop pollution
15 index (P), the target hazard quotient (THQ) (element) and the hazard index (HI) (all elements) were assessed
16 as tools to support the evaluation. Our results showed that the degree of contamination and uptake by crops in
17 the study area were: root vegetables > leaf vegetables > fruit vegetables > cereals. With the exception of corn,
18 other crops pose a significant risk to human health which is dominated by the Tl content. In addition the Cu in
19 carrot samples suggests hyper accumulation at the site and poses a high risk to human health. The results
20 provide direct evidence of significant food chain exposure and identifies the need for Tl-focused management
21 of soil/plant interaction and that strategy needs to also understand the implications for behavior of
22 co-contaminants in the area.

23 **Keywords:** Potentially Toxic Elements (PTEs); thallium mine area; crop; bioconcentration factor; health risk
24 assessment

25 **1. Introduction**

26 The effect of long-term mining frequently leads to the extensive localized pollution of soil systems by a
27 specific range of potentially toxic elements (PTEs) associated with the particular ore deposit (Chen et al. 2015).
28 The PTEs entering the soil are not biodegraded and persist and accumulate in the soil system and are potentially
29 transferred into the crops over many seasonal cycles (Adama et al. 2014). This provides a potential pathway
30 directly to the human body through the food chain (Amin et al. 2013). Excessive intake of trace elements, such
31 as copper (Cu), zinc (Zn) and nickel (Ni), which are essential for human metabolism, will cause toxic effects
32 and for non-essential elements such as cadmium (Cd), chromium (Cr) and lead (Pb), small additional intake can
33 be harmful to human health (Bermudez et al. 2011; Gall et al. 2015; Zheng et al. 2015). The impact of increased
34 exposure and ingestion has been identified to cause a variety of health problems. In our case the elements

35 thallium, arsenic, cadmium, lead, nickel, copper and zinc are of local concern, associated with specific
36 mineralization and commercially viable ore exploitation. Thallium (Tl) can cause cardiovascular disease,
37 blindness and hair loss (Wappelhorst et al. 2000; Peter et al. 2005); Arsenic (As) can cause vascular diseases,
38 neurological disorders, dermal lesions and skin cancer (Patlolla et al. 2012; Tan et al. 2016); Cd can lead to lung
39 cancer, prostatic hyperplasia, renal dysfunction and fractures (Wu et al. 2018); Pb can cause damage to
40 neuroskeletal endocrine and immune systems (Pareja-Carrera et al. 2014; Grégoire et al. 2013); Excessive
41 intake of Cu can affect liver and kidney function, the central nervous system, cause depression and even lung
42 cancer (Chen et al. 2018; Ji et al. 2013). So the question arises—at what point does ‘hazard become a significant
43 risk? Consequently an effective approach to health risk assessment in mining disrupted soil environments is of
44 real practical significance, particularly for communities utilizing land for domestic crop production.

45

46 Soil exposure (ingestion, dermal contact and inhalation of PTEs in soil), contamination of drinking water and
47 food crops are major pathways for human exposure to PTEs (Chen et al. 2018). A number of studies have
48 identified that consumption of food crops contaminated with PTEs has been demonstrated to be the major
49 route for human exposure (Bermudez et al. 2011; Ji et al. 2013; Khan et al. 2008). In addition, many health
50 risk assessment studies have focused on a single category of food which underestimates risk from the total
51 “food basket” for residents contaminated environments. For instance, Li et al. (2018) only considered
52 vegetables in the study near a large-scale Pb/Zn smelter, and found that Pb and Cd pose greatest health risk
53 and leafy vegetables tend to be more contaminated than non-leafy vegetables. Cereals were the only edible
54 crops studied by Sharma et al. (2018), who found that maize grains act as hyper-accumulators of copper
55 (BCF=30.43). Lu et al. (2018) only assessed the health risks of PTEs in rice, and found that Cd migrates more
56 significantly and poses the greatest risk to human health. The study of Tl by Jia et al.(2013) of the geochemical
57 behavior of Tl at the interface of soil and green cabbage provided evidence of high accumulation of Tl, with
58 most BCF values exceeding 1, and up to a maximum value of 11. More comprehensive studies including soil,
59 drinking water and crops, conducted for example by Luo et al. (2019) found that the highest BCF values for
60 Cd and Zn.

61 We report on a comprehensive study including leaf and root vegetables and cereals, which form the main
62 dietary components in our study area. In addition, we include assessment for a wide range of PTEs, associated
63 with the mining region. Typically As, Cd, Cr, cobalt (Co), Cu, Pb, Zn, Ni and mercury (Hg) which are the
64 almost frequently assessed in mining and industrially contaminated soils and associated crop exposure human
65 health risk assessments (Chen et al. 2018; Viraraghavan et al. 2011; Sharma et al. 2018; Lu et al. 2018; Lian et
66 al. 2019). The inclusion of Tl in these assessments is extremely rare. Even though Tl is a highly toxic element,
67 comparable to that of Cd, Hg and Pb (Peter et al. 2005; Viraraghavan et al. 2011), its general abundance is low
68 and is not a common environmental pollutant unlike those highlighted above. Consequently there are few
69 examples of field study data and little is known about its food chain transfer.

70 Thallium minerals are rare in nature, and Tl is generally found in soils with low concentrations (Xiao et
71 al., 2004). For soils, Tl is often excluded from the list of elements to be analyzed, as its concentration in
72 uncontaminated soils range from 0.01 to 3 mg/kg, while most soils contain Tl at concentrations of less than 1
73 mg/kg (Fergusson, 1990). For plants, the average concentrations of Tl in the land plants reported by Bowen
74 (1979) were 0.08-1.0 mg/kg and that in the edible plants were 0.03-0.3mg/kg, on dry weight basis. Thus, a
75 better understanding of the local environmental impact of Tl is essential to identify, seek remediation options
76 and manage any Tl-related health issues. We report on an assessment of common food basket ingredients of

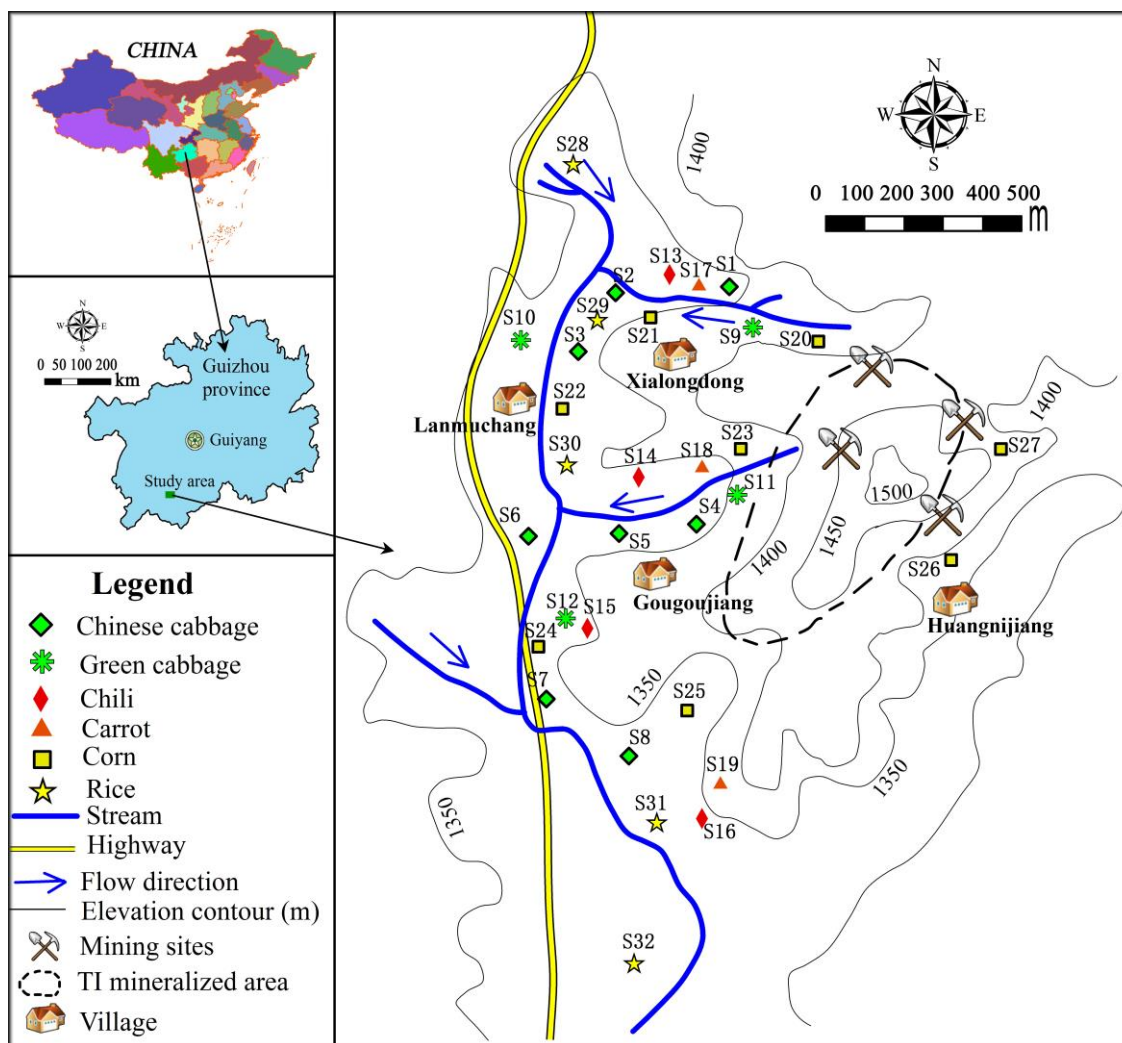
77 the population living in and around a Tl mining area. This included six common food crops (Chinese cabbage,
78 cabbage, pepper, radish, chilli, corn and rice) for which the contents of Tl, As, Cd, Pb, Ni, Cu and Zn were
79 determined and used in health risk assessment.

80

81 **2. Materials and methods**

82 *2.1. Study area*

83 The study area is located in a village in southwest Guizhou province, China (Jia et al. 2013). This area has
84 an independent thallium deposit and coal mine, where arsenic ore is also concentrated. The mining history of
85 the study area shows activity for more than 300 years (Xiao et al. 2003). Now the area has been widely
86 developed for agriculture and residential land, the majority of residents have permanent dwellings and total
87 about 1000 people. The climate of the study area is mild and humid, with an average temperature of 14°C, the
88 annual precipitation is 1300 ~ 1500 mm, with the main rainy season in summer and autumn concentrated in July
89 and August (Xiao et al. 2003). Investigations during fieldwork highlights that these residents predominantly use
90 locally grown (personal and purchased from neighbors) staple foods and the economic development of the
91 district is relatively poor. The main local food crops are: rice, corn, Chinese cabbage, green cabbage, chili and
92 carrot. Cases of more than 400 people have been found to be poisoned by thallium in the study area since the
93 1960s, which highlights extensive pollution of the region (Jia et al. 2013).



94
95 **Figure 1. Sampling sites for six crop types in thallium mining area.**

96 *2.2. Sampling and analytical methods*

97 *2.2.1. Sampling methods and preparation*

98 Six common food crop products, namely, Chinese cabbage, green cabbage, chili, carrot, corn and rice,
99 were used as samples of food exposure (Figure 1). In order to ensure representative sampling, we randomly
100 selected the edible parts of many plants at each sampling point, and combined to make a composite sample.
101 Each crop was collected at its most mature/ripe stage (within the same season/in the year 2018). The edible parts
102 of the crops were carefully cleaned of attached soil and washed with deionized water, then the surface moisture
103 was dried with filter paper, and dried in the oven for one hour (100 °C). Finally, drying to constant weight at 60
104 °C. The dried crops are ground into fine powder and pass 100 mesh nylon sieve, then stored in sealed
105 polyethylene bags at 4 °C.

106 Soil samples were collected at the crop sampling site at the same time. Sampling was to a depth of 0~
107 20cm (equivalent to general depth of crop roots in the soil), stored in polyethylene bags, and then transported
108 back to the laboratory. Soil samples dried naturally in air and obvious impurities removed, then ground into fine
109 powder and pass 100 mesh nylon sieve, stored in sealed polyethylene bags at 4 °C.

110 2.2.2. Analytical methods

111 The analytes: Tl, As, Cd, Pb, Ni, Cu and Zn were determined by inductively coupled plasma-atomic
 112 emission spectrometry (ICP-AES, JY38S, Jobin Yvon, Longumeau, France). Analytical duplicates, reagent
 113 blank, national standard soil sample (GBW07401) and plant sample (GBW07602) were used for quality control
 114 during instrument testing. Crop samples were digested using HNO₃ and H₂O₂. Soil samples were digested with
 115 aqua regia. Concentrations of PTEs in soil and crop digestion solution were determined by ICP-AES. The
 116 determination procedure is as described in our previous reports (Jiang et al. 2019). Average recoveries for PTEs
 117 plant and soil reference materials were 94.8% to 106.5% (RSD: 0.18%–3.72%), respectively.

118 2.3. Contamination assessment of crops

119 The single factor pollution index can evaluate the pollution degree of a single PTE in crops, and the
 120 comprehensive pollution index can evaluate the comprehensive pollution level of all pollutants in each crop
 121 (Li et al. 2018). The two indices can be used together and are considered to be a comprehensive and practical
 122 method in the assessment of plants contamination levels (Li et al. 2018; Yang et al. 2008), and are shown in Eqs.
 123 (1) and (2):

$$124 P_i = C_i/S_i \quad (1)$$

125 Where P_i is the single factor pollution index of pollutants in a crop; C_i is the measured value of pollutants
 126 in a crop (mg/kg, FW) on fresh weigh basis; S_i is the assessment standard for pollutants, usually the maximum
 127 allowable concentration. We used the maximum allowable concentration as an appropriate level for the
 128 population being studied due to high localized contamination. Our concern was to use data relvant to local
 129 population and then any standards provided by other organizations. The regulated pollution level of China is
 130 applied for S_i for the criteria, with GB 2762-2017 for As, Cd and Pb , GB 15199-94 for Cu and GB 13106-91 for
 131 Zn on a fresh weight basis. At present, China only sets nickel limits for margarine and other oils, but not for
 132 other foods, so we obtained the limited level of Ni in crops from FAO/WHO on a fresh weight basis
 133 (FAO/WHO, 2011). Because the concentration of thallium is very low in nature, there is no specific food limit
 134 standard for thallium at present so we set the limit level for Tl in food crops as 0.3 mg/kg on the basis of dry
 135 weight (Bowen, 1979) (more details provided in Table 2).

$$136 P = \sqrt{\frac{\overline{P}_i^2 + P_{i\max}^2}{2}} \quad (2)$$

137 Where P is the comprehensive pollution index, \overline{P}_i is the average of the single-factor pollution index, and
 138 $P_{i\max}$ is the maximum of the single-factor pollution index.

139 The evaluation of single factor pollution index and comprehensive pollution index can be classified into
 140 five grades: Class I, $P \leq 0.7$ is safe; Class II, $0.7 < P \leq 1$ is clean; Class III, $1 < P \leq 2$ is slightly polluted; Class IV,
 141 $2 < P \leq 3$ is moderately polluted; Class V, $P > 3$ is heavily polluted.

142 2.4. Bioconcentration factor

143 The crop bioconcentration factor (BCF) is the ratio of the content of the element in the plant to the content
 144 of the element in the soil of the corresponding cultivated soil. The BCF is a parameter that characterizes the
 145 effect of soil element content on food chain. It is dimensionless and so objectively reflects the potential of plant
 146 to accumulate trace elements from soil. The higher the BCF value is, the stronger the ability of the plant to
 147 accumulate the element and the weaker its ability to act as a barrier to soil metal pollution. Where $BCF > 1$ it
 148 indicates that the plant is a net accumulator of the element (Yang et al. 2008). The bioconcentration factor (BCF)
 149 was calculated using by the following equation:

$$150 \quad BCF = \frac{C_{plant}}{C_{soil}} \quad (3)$$

151 Where C_{plant} and C_{soil} are the concentrations of PTEs in crops (based on dry weight) and soil (based on dry
 152 weight) sampled from the same site, respectively (Lian et al. 2019).

153 2.5. Health risk assessment

154 The target hazard quotient (THQ) and hazard index (HI) were used to assess the potential health risks of
 155 exposure to PTEs through consumption of crops. THQ is the ratio of PTE exposure dose to reference dose. If
 156 $THQ > 1$, it indicates that the PTE has potential health risks to the exposed population, and the calculation is as
 157 follows (USEPA 1989):

$$158 \quad THQ = \frac{C \times CF \times IR \times EF \times ED}{BW \times AT \times RfD} \quad (4)$$

159 Where C is the concentration of PTEs in edible parts of food crop (on dry weight basis, mg / kg). The CF is
 160 the conversion factor from fresh to dry weight was 0.085 (Rattan et al. 2005); IR is consumption rate of a food
 161 crop, (kg / person / day; FW); EF is exposure frequency (365day/year); ED is exposure period (70year); BW is
 162 the body weight, and the average BW for adult is set to 60kg; AT is the average exposure time
 163 ($ED \times 365 \text{ day/year}$); RfD is the daily oral reference dose, RfD for Tl, As, Cd, Pb, Ni, Cu, Zn are 0.00008, 0.0003,
 164 0.001, 0.0035, 0.02, 0.04, 0.3 mg/kg/day, respectively (IRIS 2003; USEPA 2009). The average daily
 165 consumption of Chinese cabbage, carrots, green cabbage, chili, corn and rice were 0.08, 0.041, 0.029, 0.015,

166 0.002, 0.25 kg / person / day for adults, on fresh weight basis (Zhao et al. 2016; Hulshof 1998).

167 The Hazard Index (HI) indicates that the total risk of all PTEs in food crops (USEPA 1989). If $HI \leq 1$, there
168 was no significant negative effect; $HI > 1$, indicating a high probability of negative effects on human health; and
169 $HI > 10$, indicating a chronic toxic effect.

$$170 \quad HI = \sum_{n=1}^t THQ_n; \quad t=1,2,3... n \quad (5)$$

171 Where n is the nth PTE. In this study, HI is the sum of THQ of Tl, As, Cd, Pb, Ni, Cu and Zn.

172 3. Results and discussion

173 3.1 PTE in farmland soils and food crops

174 There are two main reasons for the excess of PTE in crop: the content of PTE in farmland soil is too high,
175 or the crops have a good ability to absorb one or several elements from the soil. Therefore, the information of
176 PTEs content in farmland soil can help better understand of the cause of the contamination of edible crops. The
177 concentrations for PTE in farmland soil from the thallium mine area are presented in Table 1. The average
178 concentrations of Tl, As, Cd, Pb, Ni, Cu, and Zn were 34.69 mg/kg, 156.95 mg/kg, 0.43 mg/kg, 35.00 mg/kg,
179 63.41 mg/kg, 87.00 mg/kg, and 95.34 mg/kg, respectively; corresponding to 48.72, 7.84, 0.65, 0.99, 1.62, 2.72,
180 and 0.96 times the background values of soil in Guizhou Province. It indicates that the Tl in the farmland soil
181 seriously exceeds the environmental standard, with As, Ni and Cu more than two times the standard.
182 Distribution, source identification and detailed contamination status of PTE in the soil of the study area can be
183 seen from our previous reports (Jiang et al. 2019).

184
185 **Table 1** Descriptive statistics for PTEs contamination status of the farmland soil (mg/kg).

| Element | Min | Max | Mean±S.D | B V | B V±S.D |
|---------|------|------|--------------|-------|-------------|
| Tl | 14 | 124 | 54.69±24.37 | 0.712 | 0.712±0.293 |
| As | 106 | 404 | 206.00±82.58 | 20.0 | 20.0±14.55 |
| Cd | 0.20 | 0.80 | 0.43±0.14 | 0.659 | 0.659±1.406 |
| Pb | 15 | 95 | 35.00±18.07 | 35.2 | 35.2±19.59 |
| Ni | 27 | 125 | 63.41±22.08 | 39.1 | 39.1±22.4 |
| Cu | 31 | 127 | 87.00±25.66 | 32 | 32±20.76 |
| Zn | 29 | 146 | 95.34±28.18 | 99.5 | 99.5±56.01 |

186 Min = minimum; Max = maximum; CV = coefficient of variation; S.D = standard deviation.

187 BV=Background Values, According to the Background value of soil elements in China (China National Environmental Monitoring
188 Centre, 1990)

189

190 The concentrations of PTEs and food standards for the edible parts of different crops in the study area are shown
191 in Table 2. Except for the limit levels of As, Cd, Pb, Ni, Cu and Zn, the other listed concentrations are based on
192 dry weight. The concentrations of Tl, As, Cd, Pb, Ni, Cu and Zn in the six crops ranged from 0.3~494.4 mg/kg,

193 0.01~1.39 mg/kg, 0.01~0.85mg/kg, 0.03~12.0 mg/kg, 0.1~18.4 mg/kg, 0.2~6035 mg/kg, 14~119 mg/kg,
194 respectively with the concentration of PTEs in vegetables (Chinese cabbage, green cabbage, chili, carrot) being
195 higher than that in grains (corn, rice). If data are presented on a fresh weight basis, the concentration of As, Cd
196 and Pb are close to element limit levels for individual crops. Except for Cu in carrot, the concentrations of Ni,
197 Cu and Zn in crops are very low. The concentration of Tl was relatively high in all crops, especially in green
198 cabbage, it indicates that Tl is easily taken up by plants and Green cabbage act as hyper-accumulators of Cu,
199 which has been identified by many studies (LaCoste et al. 2001; Xiao et al. 2004; Al-Najar et al. 2005;
200 Pavlickova et al. 2005; Jia et al. 2013). Thus, it is not surprising that the concentration of Tl in Green cabbage
201 is very high in this study, due to the dual factors of elevated Tl in the soils and the super-accumulation property
202 of Green cabbage.

203 Of the seven PTEs, Tl is highlighted as the main pollutant in crops, and the average concentration of Tl in
204 all crops is significantly higher than the relevant standard. Compared with the previous reports of PTEs
205 (especially Tl) in vegetables and crops in this area by Prof. XIAO Tangfu's group, the content of PTEs in
206 many crops is very similar or slightly (but not significantly) lower. It can be inferred that the concentration of
207 PTEs in crops is species-specific and there is variation with the concentration of PTEs in soil where the crops
208 grow. The content of PTEs in farmland soil will migrate to crops, which will lead to the decrease of PTEs in
209 soil and crops year by year. However, for strongly accumulating crops depletion of PTEs in soil by crop
210 growth is too low to achieve sanitation of heavily polluted sites (sager, 1988). On the other hand, PTEs in the
211 farmland soil are influenced by agricultural activities which include the addition of wastewater sludge and
212 cinders from local residents as soil improvers, in addition to inputs from continued mining activities (Jiang et
213 al. 2019). Therefore, a number of factors may have led to changes in uptake of PTEs over time.

214 Previous studies have shown that plants with thick roots and strong active transport ability can absorb
215 more PTEs from the soil (Shahid et al. 2016; Bondada et al. 2004). In this study, the crops with the highest As,
216 Cd, Ni, Cu and Zn contents is carrot, which is consistent with other research (Shahid et al. 2016; Bondada et al.
217 2004). In addition to the absorption of PTEs by the root system (from the soil), the absorption of the leaves
218 (from the atmosphere) is also the main way to transfer PTEs to crops (Shahid et al. 2016). This means that
219 Chinese cabbage, green cabbage, chili and carrot are likely to be the more significant accumulators than cereals.

220 Therefore compared with other crops, the very high concentration of Cu in carrot and Tl in green cabbage
221 is not surprising but also it can be clearly seen that the uptake of PTEs by crops varies both with the PTEs and
222 crop types, due to the different physiological and absorption mechanisms.

223

Table 2. Concentrations of PTE in various crop (DW, mg/kg except where indicated).

| Sample | Statistics | Tl | As | Cd | Pb | Ni | Cu | Zn |
|-----------------------|-------------|----------|------------|-----------|-----------|-----------|-----------|---------|
| Chinese cabbage (n=8) | Mean±SD | 2.1±2.0 | 0.76±0.31 | 0.27±0.10 | 3.1±3.7 | 2.1±0.7 | 6.0±1.6 | 54±19 |
| | Min | 0.7 | 0.17 | 0.15 | 0.4 | 1.0 | 3.5 | 30 |
| | Max | 5.4 | 1.25 | 0.43 | 12.0 | 3.3 | 9.2 | 92 |
| | Limit level | 0.3 | 0.5 | 0.2 | 0.3 | 67 | 10 | 20 |
| Green cabbage (n=4) | Mean±SD | 261±221 | 0.77±0.27 | 0.30±0.09 | 0.7±0.2 | 1.9±1.1 | 3.1±0.5 | 51±28 |
| | Min | 31.7 | 0.38 | 0.17 | 0.5 | 0.6 | 2.7 | 32 |
| | Max | 494.4 | 0.98 | 0.38 | 1.0 | 2.7 | 3.8 | 92 |
| | Limit level | 0.3 | 0.5 | 0.2 | 0.3 | 67 | 10 | 20 |
| Chili (n=4) | Mean±SD | 3.9±1.0 | 0.25±0.03 | 0.06±0.03 | 0.22±0.03 | 1.9±0.5 | 8.0±0.7 | 18±3.5 |
| | Min | 2.9 | 0.22 | 0.04 | 0.18 | 1.2 | 7.1 | 14 |
| | Max | 5.3 | 0.29 | 0.10 | 0.25 | 2.5 | 8.6 | 21 |
| | Limit level | 0.3 | 0.5 | 0.05 | 0.1 | 67 | 10 | 20 |
| Carrot (n=3) | Mean±SD | 21.7±5.5 | 1.24±0.16 | 0.80±0.06 | 1.86±0.20 | 16.8±1.65 | 5275±695 | 113±6.5 |
| | Min | 16.3 | 1.07 | 0.73 | 1.64 | 15.1 | 4671 | 106 |
| | Max | 27.2 | 1.39 | 0.85 | 2.01 | 18.4 | 6035 | 119 |
| | Limit level | 0.3 | 0.5 | 0.1 | 0.1 | 67 | 10 | 20 |
| Corn (n=8) | Mean±SD | 1.5±0.9 | 0.02±0.004 | 0.04±0.06 | 0.05±0.03 | 0.22±0.08 | 1.7±0.7 | 19±2.3 |
| | Min | 0.78 | 0.01 | 0.01 | 0.03 | 0.1 | 1.4 | 17 |
| | Max | 3.08 | 0.03 | 0.16 | 0.11 | 0.4 | 3.1 | 23 |
| | Limit level | 0.3 | 0.5 | 0.1 | 0.2 | 67 | 10 | 20 |
| Rice (n=5) | Mean±SD | 2.0±1.9 | 0.16±0.41 | 0.02±0.01 | 0.05±0.01 | 0.12±0.05 | 0.74±0.54 | 18±3.7 |
| | Min | 0.3 | 0.10 | 0.01 | 0.03 | 0.1 | 0.2 | 14 |
| | Max | 5.2 | 0.21 | 0.03 | 0.06 | 0.2 | 1.6 | 24 |
| | Limit level | 0.3 | 0.2 | 0.2 | 0.2 | 67 | 10 | 50 |

224 Limit level of Tl was obtained from Bowen et al (Bowen, 1979); limit level of As, Cd and Pb were obtained from GB 2762-2017
 225 (MHPRC, 2017); limit level of Cu obtained from GB 15199-94; limit level of Zn obtained from GB 13106-91; limit level of Ni was
 226 obtained from AO/WHO (FAO/WHO, 2011); concentrations of limit level in GB 2762-2017, GB 15199-94, GB 13106-91 and
 227 FAO/WHO are based on fresh weight.

228 3.2. Bioaccumulation of PTEs

229 The BCF values of PTEs in different crops are shown in Table 3. The BCF values for Cd and Zn is higher
 230 than that of other elements in most crops, following the order of Cd > Zn > Cu > Tl > Ni > Pb > As, confirming
 231 behavior of the elements seen elsewhere (Lu et al. 2018; Lian et al. 2019), which also indicates that Cd and Zn
 232 have high transfer from soils to food crops.

233 In general, the order of BCF in crops is: rhizome vegetables (carrot) > leafy vegetables (Chinese cabbage
 234 and green cabbage) > fruit vegetables (chili) > cereal (rice and corn). This is consistent with the results of Li et
 235 al. (2018). Li et al. (2018) and Lian et al. (2019) also indicates that the cause of this phenomenon is related to
 236 strong root absorption and atmospheric deposition (leaf absorption).

237 The bioaccumulation of Cu in carrot was very strong (BCF=83.744), followed by Cd (BCF=1.932) and Zn
 238 (BCF=0.977). The bioaccumulation of Tl in green cabbage Tl was very strong (BCF=4.994). The highest

239 enrichment for Pb is found in Chinese cabbage (0.100).The different accumulation across elements and crop
240 types is likely to be more strongly controlled by physiological requirements (Shahid et al. 2016).

241 In general, it is worth noting here that several crops have strong bio-enrichment capacity for Cu, Tl, Cd
242 and Zn. Whilst Cd and Zn are well known to be more mobile in soil and it is not be surprising to see high
243 BCFs. The strong bio-enrichment of Tl in Chinese cabbage has been discussed above. Previous studies have
244 also reported high Cu uptake by crops. For example, Lian et al. (2019) reported that the concentration of Cu in
245 roots is higher than other parts for most vegetables. The pollution pathways of PTEs to crops include root
246 absorption (from farmland soil) and leaf absorption (from atmospheric deposition) (Shahid et al., 2016).
247 Besides, crops with massive roots can absorb more PTEs from the soil (Shahid et al., 2016). Thus, the leaf of
248 carrot can absorb more PTEs from the atmosphere due to the larger leaf surface; and the large root/tuber of
249 carrots may absorb more PTEs from the soil. On the other hand, our previous work confirmed that soil Cu in
250 the study area is only slightly polluted, and the waste gas and slag produced by coal combustion are the main
251 sources of Cu pollution (Jiang et al. 2019). Thus, the high concentration of Cu and its BCF in carrots may be
252 related to both soil absorption and atmospheric deposition, which is consistent with Li et al.(2018).

253

254

Table 3. Bio-concentration factors (BCF) for PTEs in different crops.

| Crop | Tl | As | Cd | Pb | Ni | Cu | Zn |
|-----------------|-------|-------|-------|-------|-------|--------|-------|
| Chinese cabbage | 0.048 | 0.005 | 0.661 | 0.100 | 0.034 | 0.070 | 0.542 |
| Green cabbage | 4.994 | 0.004 | 0.760 | 0.018 | 0.039 | 0.038 | 0.575 |
| Chili | 0.068 | 0.002 | 0.152 | 0.008 | 0.030 | 0.105 | 0.194 |
| Carrot | 0.366 | 0.006 | 1.932 | 0.062 | 0.397 | 83.744 | 0.977 |
| Corn | 0.028 | 0.001 | 0.101 | 0.002 | 0.004 | 0.029 | 0.253 |
| Rice | 0.048 | 0.001 | 0.056 | 0.002 | 0.002 | 0.007 | 0.284 |

255

256 3.3. Significance of Contamination of the edible crops

257 The contamination by Tl is very serious in all crops (Class V (heavily polluted)), especially in the case of
258 green cabbage, (Table 4). The pollution indexes for As, Cd, Ni and Zn in very crops are low (all Class I (safe)).
259 For As, Cd, Pb, Ni, Cu and Zn in cereals (corn and rice) classification is Class I (safe). The for Cu in carrot
260 classification was very high (Class V: heavily polluted), followed by Pb (Class III: slightly polluted) The
261 contamination level of As, Cd, Ni, Cu and Zn in Chinese cabbage, cabbage and chili showed Class I, the
262 pollution index of were all less than 0.7 (safe) in these three vegetables. There is no Pb contamination in
263 Chinese cabbage (Class II: clean), and in carrot was slightly polluted (Class III).

264 Except for the specific pollution in some crops, the order of degree of PTE contamination is $Tl > Pb > As >$
 265 $Cd > Zn > Cu > Ni$ and for crops pollution follows: rhizome vegetable > leaf vegetable > fruit vegetable > cereal,
 266 as for the bioaccumulation factor for PTEs.

267

268 Table 4. Single-factor (P_i) and comprehensive pollution (Class) indices for the edible crops.

| elements | Chinese cabbage (n=8) | | Green cabbage (n=4) | | Chili (n=4) | | Carrot (n=3) | | Corn (n=8) | | Rice (n=5) | |
|----------|-----------------------|-------|---------------------|-------|-------------|-------|--------------|-------|------------|-------|------------|-------|
| | P_i | Class | P_i | Class | P_i | Class | P_i | Class | P_i | Class | P_i | Class |
| Tl | 7.017 | V | 870.583 | V | 12.858 | V | 72.333 | V | 4.800 | V | 6.633 | V |
| As | 0.129 | I | 0.131 | I | 0.043 | I | 0.211 | I | 0.003 | I | 0.067 | I |
| Cd | 0.115 | I | 0.125 | I | 0.102 | I | 0.680 | I | 0.029 | I | 0.09 | I |
| Pb | 0.864 | II | 0.205 | I | 0.183 | I | 1.581 | III | 0.020 | I | 0.020 | I |
| Ni | 0.003 | I | 0.002 | I | 0.002 | I | 0.021 | I | 0.001 | I | 0.001 | I |
| Cu | 0.051 | I | 0.027 | I | 0.068 | I | 44.843 | V | 0.015 | I | 0.006 | I |
| Zn | 0.230 | I | 0.216 | I | 0.074 | I | 0.479 | I | 0.082 | I | 0.031 | I |
| P^* | 5.033 | V | 624.855 | V | 9.191 | V | 52.567 | V | 3.430 | V | 4.739 | V |

269 P^* : Comprehensive pollution index

270 Note: Class I, safe; Class II, clean; Class III, slightly polluted; Class IV, moderately polluted; Class V, heavily polluted.

271

272 According to the comprehensive pollution index (P), the rank order for crop pollution was Green
 273 cabbage > Carrot > Chili > Chinese cabbage > Rice > Corn,. In this sequence, the comprehensive pollution
 274 levels of the six crops all reached Class V (heavily polluted), and the pollution of Green cabbage is extremely
 275 serious.

276 3.4 Human health risk

277 The THQ values of each element studied and the HI values of each crop are shown in Table 5. The order of
 278 health hazards to human in most crops was $Tl > As > Pb > Cd > Cu > Zn > Ni$, which is similar to their
 279 single-factor pollution index (P_i) in crops. Therefore, the higher the pollution index of crops, the more adverse
 280 the trend will be to human health. Hazard index (HI) values for each crop are ordered as $Tl > Cu > As > Pb >$
 281 $Cd > Zn > Ni$. The THQ value of Tl in all crops except corn exceeded the safety threshold of 1.0, and long-term
 282 consumption of these crops would cause harm to the body. Only the THQ of Cu in carrot exceeded the safety
 283 threshold (1.0), and consumption of Cu from other crops does not pose a risk to humans. The majority of PTEs
 284 studied (As, Cd, Pb, Ni and Zn) do not pose significant risk of harm to humans in all crops ($THQ \leq 1$). It can be
 285 seen that even if some elements exceed the standard in crops or polluted in crops (see Table 1, 4), they do not
 286 pose a health hazard (e.g., Pb in carrot). In addition, Excessive PTE in the soil does not pose a risk to human
 287 health due to its low bioenrichment capacity through the food chain exposure.(e.g., As and Ni).

288 Figure 2 and Table 5 confirm that the exposure to Tl in the crops is the most harmful (except for Cu in
 289 carrots). The order of decreasing HI was Green cabbage > Carrot > Rice > Chinese cabbage > Chili > Corn.
 290 Only corn demonstrated no negative effect on human health. Green cabbage was far more significant for human
 291 health than other crops, followed by carrot. The long-term consumption of rice, Chinese cabbage and chili is of
 292 concern from results here, while consumption of green cabbage and carrots will cause chronic toxic effects on
 293 human body ($HI > 20$).

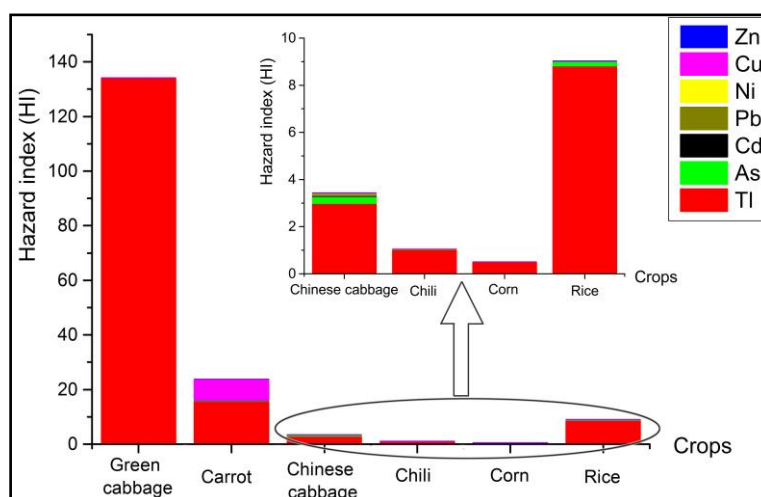
294 Table 5. THQ and HI values of PTE for local residents via crop consumption in the thallium mine area.

| Crops | Target hazard quotient (THQ) | | | | | | | H I |
|-----------------|------------------------------|-------|-------|-------|-------|-------|-------|---------|
| | Tl | As | Cd | Pb | Ni | Cu | Zn | |
| Chinese cabbage | 2.982 | 0.287 | 0.031 | 0.099 | 0.012 | 0.017 | 0.020 | 3.448 |
| Green cabbage | 134.124 | 0.011 | 0.012 | 0.009 | 0.004 | 0.003 | 0.007 | 134.170 |
| Chili | 1.025 | 0.018 | 0.001 | 0.001 | 0.002 | 0.004 | 0.001 | 1.052 |
| Carrot | 15.755 | 0.240 | 0.046 | 0.031 | 0.049 | 7.661 | 0.022 | 23.804 |
| Corn | 0.510 | 0.002 | 0.001 | 0.001 | 0.001 | 0.001 | 0.002 | 0.518 |
| Rice | 8.810 | 0.187 | 0.008 | 0.005 | 0.002 | 0.007 | 0.021 | 9.040 |
| Total | 157.206 | 0.745 | 0.099 | 0.146 | 0.070 | 7.693 | 0.073 | |

295

296 This study suggests that local residents may be over-exposed to Tl and Cu, through crops alone. However,
 297 exposure to PTEs through other exposure pathways such as drinking water, soil (ingestion, inhalation, and
 298 dermal contact) further increases the health risk from exposure to PTEs (Zheng et al. 2007; Järup 2003; Simsek
 299 et al. 2000). It is worth noting that although the degree of pollution and bioconcentration of rice is lower than
 300 that of Chinese cabbage and chili, the risk to human health is greater than that from consumption of Chinese
 301 cabbage and chili due to significance of rice in the daily diet of residents in the study area.

302



303 **Fig. 2. Hazard index for consumption of subsistence crops grown in thallium mining area.**

304 **4. Conclusions**

305 In this study, we compared the pollution status of a number of PTEs in crops from a thallium mining area,
306 analyzed the absorption ability of PTEs in vegetables and cereals, and evaluated the health risks caused by
307 eating these crops. It was found that Tl intake via consumption of all food crops posed high risk to human health
308 from the study area as compared to other elements, due to its high toxicity and high concentration in crops. It
309 was also found that Green cabbage was found to be hyper-accumulator of Tl, whereas, carrot acted as
310 hyper-accumulator of Cu. This could be attributed to variable physiological requirements and uptake
311 mechanisms of different crops and PTEs status in the environment. Moreover, the contamination of PTEs in
312 crops, either through the uptake from root absorption (from farmland soil) or leaf absorption (from atmospheric
313 deposition), poses a health concern to local residents. In addition, food crops should be avoided in the study area,
314 and plants with high uptake of PTEs should be planted to reduce the concentration of PTEs (especially thallium)
315 in food crops. We also highlight the need to use location specific data – particularly on consumption patterns to
316 allow realistic risk analysis.

317

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320

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